

1 **Title**

2 The invasive red swamp crayfish (*Procambarus clarkii*) increases infection of the amphibian
3 chytrid fungus (*Batrachochytrium dendrobatidis*)

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21 **Abstract**

22 Emerging infectious diseases are increasingly recognized as a severe threat to wildlife.
23 Chytridiomycosis, caused by *Batrachochytrium dendrobatidis* (Bd), is considered one of the most
24 important causes for the decline of amphibian populations worldwide. Identifying potential
25 biological reservoirs and characterizing the role they can play in pathogen maintenance is not only
26 important from a scientific point of view, but also relevant from an applied perspective (e.g. disease

control strategies), especially when worldwide distributed invasive species are involved. We aimed (1) to analyse the prevalence and infection intensity of Bd in the invasive red swamp crayfish (*Procambarus clarkii*) across the western Andalusian region in Spain; and (2) to assess whether the presence of crayfish affects the prevalence and infection intensity of Bd in amphibians of Doñana Natural Space (DNS), a localized, highly protected area within the Andalusian region. First, we found that infection prevalence in crayfish guts was 1.5% regionally (four out of 267 crayfish were qPCR positive to Bd, all of them belonging to the same Andalusian population); qPCR positives were histologically confirmed by finding zoosporangia of Bd in gastrointestinal walls of the red swamp crayfish. Second, we found a higher prevalence of Bd infection in DNS (19% for crayfish and 28% for amphibians on average), a place with great diversity and abundance of amphibians. Our analyses showed that prevalence of Bd in amphibians was related to the presence of the red swamp crayfish, indicating that this crayfish could be a suitable predictor of Bd infection in co-occurring amphibians. These results suggest that the red swamp crayfish might be a possible reservoir for Bd, representing an additional indirect impact on amphibians, a role that had not been previously recognised in its invasive range.

42

43 **Keywords**

44 Amphibians, chytridiomycosis, Doñana, emerging infectious diseases, histological analysis, non-
45 amphibian host, transmission.

46

47 **Introduction**

48 Globalized infectious diseases, particularly fungal pathogens, are one of the greatest threats
49 to wildlife (Fisher et al. 2012). Amphibians are the most threatened vertebrates globally, being
50 highly susceptible to climatic change, habitat fragmentation and overexploitation (Stuart et al.
51 2004). Moreover, they are affected by the most devastating panzootic so far, amphibian
52 chytridiomycosis, which is closely related to amphibian declines (Voyles et al. 2009; GISD 2018;

53 Scheele et al. 2019). The amphibian-killing pathogen, *Batrachochytrium dendrobatidis* (hereafter
54 “Bd”), infects keratinizing epithelial cells in amphibian skin (Berger et al. 1998; van Rooij et al.
55 2015), causing osmotic damage and often death (Daszak et al. 2003; Voyles et al. 2009). Bd is
56 considered as one of the 100 worst invasive species worldwide (GISD 2018), whose origin has been
57 recently attributed to Asia, and the pathways of its spread at the global scale are mainly linked to
58 amphibian pet trade (O’Hanlon et al. 2018). Global trade is also resulting in the evolution of further
59 hypervirulent fungal lineages across a diverse range of host species and biomes (Farrer et al. 2011).

60 Because pathogenicity of Bd is temperature-dependent, prevalence of Bd is usually higher in
61 early spring and lower in late summer or autumn (Kriger and Hero 2007), which affects the
62 temporal patterns of exposure (Walker et al. 2010; Xie et al. 2016). The highest rates of Bd-
63 infection often coincide with the breeding season of many amphibian species during winter, making
64 them more vulnerable (Gervasi et al. 2017). However, viability of the Bd pathogen decreases
65 through the seasons in areas where water temperature rises over 30°C and waterbodies dry up
66 during the dry season (Piotrowski et al. 2004, Doddington et al. 2013). Thus, amphibians with a
67 highly variable response to different levels of infection can be reservoirs of Bd, maintaining this
68 pathogen in the environment (Woolhouse et al. 2001; Gervasi et al. 2017; Brannelly et al. 2018).
69 Others have hypothesized on the saprophytic feeding of Bd, which is able to live without an
70 amphibian host (Speare et al. 2001). In addition, in spite of being an amphibian fungal disease, there
71 is increasing interest in evaluating the potential role of non-amphibian species, which co-occur in
72 aquatic ecosystems, as reservoirs or carriers of chytridiomycosis (van Rooij et al. 2015).
73 Identification of potential reservoirs and characterization of the role they play in Bd maintenance
74 and disease dynamics is not only important from an evolutionary perspective, but can also provide
75 important insights on disease control strategies.

76 The red swamp crayfish (*Procambarus clarkii*), native to north-eastern Mexico and south-
77 central US, has been intentionally introduced worldwide, becoming the most widespread crayfish in
78 the world (Oficialdegui et al. 2019). Over the last 45 years, southern European freshwaters have

79 been widely invaded (Kouba et al. 2014), causing, among others, dramatic declines of many
80 amphibian populations and negatively impacting their community structure (Ficetola et al. 2012),
81 with an important decrease in amphibian richness (Cruz et al. 2006). The recent finding that the red
82 swamp crayfish and other crayfish species can become infected by Bd in their native range
83 (Louisiana, US), and potentially transmit infection to amphibians (McMahon et al. 2013, Brannelly
84 et al. 2015a, but see Brannelly et al. 2015b), suggests the need for evaluating the role of invasive
85 crayfish as carriers of the amphibian fungus in its invaded range. This is especially important in
86 places where crayfish species are particularly prolific and colonize a wide range of aquatic
87 environments where amphibians occur (GISD 2018).

88 The red swamp crayfish is especially abundant in the southern part of the Iberian Peninsula,
89 including the Doñana Natural Space (DNS), a protected area located in western Andalusia where
90 the conservation value of ponds for amphibians and other aquatic fauna and flora has been
91 highlighted by many studies (Gómez-Rodríguez 2007; Díaz-Paniagua et al. 2010). Due to its
92 commercial value, humans have intentionally translocated the red swamp crayfish among water
93 bodies, promoting its rapid colonization across Europe (Oficialdegui et al. 2019). Thus, studies
94 evaluating the role of the red swamp crayfish as potential reservoir in this area are of particular
95 interest, as crayfish may rapidly disperse Bd over long distances and contribute to the global Bd
96 pandemic in Europe. We conducted a study to explore these questions that comprised two sections.
97 The first one was a regional survey across the western Andalusian region, with the goal of assessing
98 the prevalence of infection and infection intensity of Bd in the red swamp crayfish in Andalusian
99 streams and ponds (i.e., crayfish infection survey). The second one was a specific sampling in a
100 localized, highly protected area within the Andalusia region (DNS), which is particularly rich in
101 amphibian populations where Bd had been circulating through the environment (Hidalgo et al.
102 2012); our aim was determining whether Bd infection prevalence and intensity in amphibians was
103 positively related to the presence of the red swamp crayfish in ponds (i.e., crayfish/amphibian
104 interaction).

105

106 **Methods**

107 *Study area*

108 In the crayfish infection survey, we sampled 11 streams of the provinces of Seville, Cadiz
109 and Huelva, in western Andalusia (South-western Spain) and three ponds within Doñana Natural
110 Space (DNS) (Fig. 1A, see Table 1). The region is dominated by the Guadalquivir basin and
111 characterized by a Mediterranean climate, with a long dry summer season, mean annual
112 temperatures between <10°C and 18°C, and mean altitude of 200 m asl. Most streams in this
113 Mediterranean region are intermittent, where streams have little flow during wet season water
114 remains in puddles along the stream during dry season. Subsequently, for the study of
115 crayfish/amphibian interaction, we sampled six different ponds in the Doñana Natural Space (DNS),
116 a localized protected area within western Andalusia which is particularly important for amphibian
117 breeding and conservation, in order to analyse the prevalence and infection intensity of Bd in
118 relation to the presence of the red swamp crayfish (Fig. 1B). The DNS, situated in the mouth of the
119 Guadalquivir River, comprises the Doñana Natural Park and the Doñana National Park, and is
120 considered one of the largest and most important wetlands of Europe. It has been declared a
121 UNESCO Biosphere Reserve, a Ramsar site, a Natural World Heritage Site, and part of the Natura
122 2000 network (see e.g. García-Novo and Marín-Cabrera 2005). As other very scarce Mediterranean
123 temporary ponds, this system of more than 3,000 water bodies that integrate DNS is considered a
124 priority habitat under the European Union Habitats Directive (Code 3170: European Commission
125 DG Environment, 2007).

126

127 *Sample collection*

128 To address the first objective, adults of the red swamp crayfish were collected during late
129 spring and early summer of 2015 from 14 sites distributed across the western Andalusian region in
130 Spain (Fig. 1A, Table 1). For most sampling sites, 20 crayfish were captured using fyke nets (Table

1), subsequently euthanized by decapitation, and their gastrointestinal (GI) tracts were removed. Samples were fixed in 70% ethanol and maintained in the refrigerator until laboratory analysis. For the second objective, sampling was performed in six ponds of DNS between early March and early April of 2018 (Fig. 1B, Table 2 and 3). While crayfish were processed as for the first objective, amphibians were captured by hand nets and skin swabs were collected by gently rubbing a sterile cotton-tipped swab along their entire body. Subsequently, swabs were kept in Eppendorf vials with a drop of pure ethanol and maintained in the refrigerator until qPCR analysis.

Real-time PCR TaqMan assay for Bd-quantification

In the laboratory, to ensure no contamination between samples, we used disposable blades to dissect the GI tract from each crayfish. In order to avoid contamination by amphibian tissue inside the GI tissue of crayfish (i.e. in case crayfish had fed on infected amphibians), the sample was washed twice with deionized pressure water. By using PrepMan Ultra Sample Preparation Reagent (Applied Biosystems), we extracted DNA from a small part of the GI tract of crayfish, and from swabs for amphibians. We quantified Bd DNA from GI tracts of crayfish and swabs using standard real-time Polymerase Chain Reaction (qPCR) procedures (Boyle et al. 2004). Amplifications were carried out in a 15µl volume reaction, which included a Bd-specific TaqMan probe (Chytr MGB) for the quantification of zoospore equivalents (ZE). We included amplification standards of 0.1, 1, 10 and 100 zoospore equivalents (ZE, where one ZE is equivalent to a single zoospore) prepared from an isolate of known cell density (IA042, Spain) and a negative control in each plate. All samples, diluted (1/10), standards and the negative control were analysed in duplicate. We considered Bd positive samples if both replicates resulted positive and the amplification curves had the expected sigmoidal shape (otherwise, a 1/100 dilution was made to prevent inhibition problems). In case of contradictory results, the sample was repeated a third time. Infection intensity was reported in zoospore equivalents (ZE) and log-transformed values to show results in graphs.

156 We considered ZE values of 0.1 (Cycle Threshold, CT < 37.0) or higher as positive for infection.
157 Infection intensities are reported as the mean and standard error, unless otherwise noted.

158

159 *Histological analyses*

160 Only crayfish samples that were determined Bd positive by qPCR in the first objective
161 (crayfish infection survey) were confirmed histologically. Samples were fixed in neutral buffered
162 10% formalin, embedded in paraffin and sectioned, before being stained with haematoxylin and
163 eosin using routine methods (Drury & Wallington 1980).

164

165 *Statistical analyses*

166 While the prevalence of Bd⁺ and infection intensity was only calculated in crayfish for the
167 crayfish infection survey, we calculated the prevalence of Bd⁺ and infection intensity in both
168 crayfish and amphibians for the crayfish/amphibian interaction. For both objectives, we calculated
169 prevalence as the proportion of Bd⁺ with respect to the total number of sampled individuals
170 (crayfish and amphibians), and infection intensity was calculated as Bd load [$\log_{10}(\text{Bd load} + 1)$] in
171 each sample. In our models, we aimed to analyse whether the presence of crayfish could have an
172 effect on the prevalence and infection intensity of Bd in amphibians. To do that, we pooled all
173 amphibian species (see Table 3 for species sampled) and both sexes to increase sample size. As we
174 captured sub-adults (non-mature phase) and adults, we used life stage as independent variable to
175 control for different time of permanence in the water. Thus, we analysed the effects of
176 presence/absence of crayfish, pond and amphibians' life stage (sub-adults and adults) on the
177 prevalence of infection and infection intensity of amphibians. We used a Generalized Linear Model
178 (GLM) with a binomial distribution to analyse the prevalence of infection and, in a second analysis
179 only with Bd⁺-amphibians, we used a linear model to analyse the infection intensity in amphibians
180 from ponds in the localized protected area (DNS). The factor “pond” was nested within the factor
181 “presence of crayfish in each pond”; both factors were included in the model because one level of

182 the factor “pond” could not combine presence and absence of crayfish at the same time but ponds
183 could differ in infection intensity. Normality of the residuals of both models was visually inspected
184 and they did not differ from normality. All analyses were performed with JMP 12 software (SAS
185 Inc.).

186

187 **Results**

188 *Crayfish infection survey*

189 Only four crayfish out of 267 were positive to Bd, which means a total prevalence of 1.5%
190 (95% CI, 0.4 – 3.8). All Bd⁺-individuals were found in one population, Olivargas stream (site 9,
191 Fig. 1A), where the prevalence reached 20% (95% CI, 5.7 – 43.7) and average infection intensity
192 was 4.1 ± 1.2 ZE (Table 1). After detecting qPCR Bd⁺, histological analyses confirmed the presence
193 of zoosporangia in all four infected GI walls of the red swamp crayfish (Fig. 2).

194

195 *Crayfish/amphibian interaction*

196 We sampled six different ponds (Fig. 1B) and captured a total of 37 crayfish and 165
197 amphibians of seven different species. A total of seven crayfish were qPCR-positive for Bd
198 infection in their GI walls, which means a prevalence of 18.9% (95% CI, 8.0 – 35.2), and average
199 infection intensity was 6.0 ± 2.0 ranging from 0.6 to 13.5 ZE. On the other hand, a total of 46
200 amphibians of three species (*Triturus pygmaeus*, *Pleurodeles waltl* and *Pelophylax perezi*) were
201 qPCR positive for Bd infection (Table 3), reaching an average prevalence of 27.6% (95% CI, 21.2 –
202 35.4) and average infection intensity was 1.9 ± 0.6 ranging from 0.1 to 68.0 ZE (Table 2). Except
203 for *P. perezi*, the other two species were widely represented in our samples. We rarely found
204 positives in anurans (one individual of *P. perezi*) with prevalence of infection of 2.4% (95% CI, 0.1
205 – 12.6) on average. However, the prevalence of infection in urodelans was of 36.6 % (95% CI, 28.1
206 – 45.7), being the prevalence within species of 35.1% (95% CI, 24.5 – 46.8) for *T. pygmaeus* and
207 39.1% (95% CI, 28.1 – 45.7) for *P. waltl* (Table 3).

208 Our generalized linear model (binomial distribution) explained a 20.6% of total variance
209 (generalized adjusted R^2) for the prevalence of infection. The prevalence of Bd in amphibians was
210 affected by the presence of crayfish ($\chi^2 = 28.5$, $p < 0.0001$), the pond ($\chi^2 = 8.8$, $p = 0.0121$) and life
211 stage of amphibians ($\chi^2 = 4.8$, $p = 0.0280$) (Fig. 3A). However, the Bd infection intensity on
212 amphibians was significantly affected by pond ($F = 2.90$, $df = 4.45$, $p = 0.0342$), but not by the
213 presence of crayfish in the pond ($F = 1.83$, $df = 1.45$, $p = 0.1838$) or life stage ($F = 1.23$, $df = 1.45$, p
214 $= 0.2737$) (Fig. 3B).

215

216 Discussion

217 Identification of non-amphibian hosts is crucial to understand the virulence, distribution,
218 spread and persistence of Bd in aquatic ecosystems worldwide. McMahon et al. (2013) observed
219 encystment of Bd within GI tracts of the red swamp crayfish from Louisiana, the native area of this
220 species. Importantly, our study confirmed the histological evidence of Bd embedded within GI
221 tracts of the red swamp crayfish in its invaded range, indicating that the red swamp crayfish could
222 be a potential carrier of this disease wherever it invades. The presence of the red swamp crayfish
223 could thus imply, besides a predatory effect on amphibians, an indirect effect through the
224 transmission of Bd, promoting the decline of amphibians in Europe or elsewhere. Our results
225 indicated that the prevalence of Bd-infection in amphibians was high when the red swamp crayfish
226 co-occurred in the same pond and, to a lesser extent, showed effects of pond (site) and amphibian
227 life stage, being the sub-adult stage more susceptible to be Bd positive. For infection intensity we
228 only found a slight effect of pond but other non-tested variables (e.g. environmental variables or
229 amphibian densities) could alter the infection intensity in amphibians. Therefore, our results suggest
230 that non-amphibian species, and the presence of red swamp crayfish in particular, could play a
231 crucial role in maintaining and spreading this emerging infectious disease (chytridiomycosis) in
232 amphibians.

233 Due to its great impact on amphibians, chytridiomycosis is considered the worst infectious
234 disease in vertebrates (Gascon et al. 2007; GISD 2018). Sudden high mortality rates in several
235 amphibian species have been related to this emerging panzootic infectious disease in the last
236 decades (Berger et al. 1998; Bosch et al. 2001; Daszak et al. 2003; Bosch and Martínez-Solano
237 2006; Skerrat et al. 2007; Scheele et al. 2019). In Andalusian streams, we found a prevalence of Bd
238 of 1.5% in wild red swamp crayfish, which was similar to that found in a previous study in its
239 native area (3.3% in the wild during spring) (Brannelly et al. 2015a). On the other hand, the red
240 swamp crayfish in the localized highly protected area (DNS) showed a total prevalence of 18.9%
241 (seven out of 37 sampled crayfish). Although our sampling size was too low to draw firm
242 conclusions, it seems to exist a tendency to a higher prevalence of Bd in DNS compared to those
243 from waterbodies across western Andalusia. It is well known that the infection of Bd is context-
244 dependent on the amphibian host (Scheele et al. 2017), the fungal virulence (Fisher et al. 2009), and
245 environmental determinants such as temperature, altitude, seasonality and/or rainfall (Berger et al.
246 2004; Kriger and Hero 2007; Walker et al. 2010; Doddington et al. 2013; Ruggeri et al. 2018). As
247 presence of Bd is higher at low temperatures (Berger et al. 2004), spatial differences in Bd
248 prevalence found between sampled areas (streams regionally and localized protected area) could be
249 rather due to the sampling timing because crayfish across western Andalusia were sampled in late
250 spring whereas those in DNS were sampled in early spring, the latter favouring Bd positives. In
251 fact, all crayfish sampled within DNS in late spring (when water temperature can exceeded 25 °C)
252 were Bd negatives. But also, the great diversity and abundance of amphibian species in DNS (Díaz-
253 Paniagua et al. 2006) could facilitate that this disease remains in the environment over time because
254 amphibians vary greatly in their susceptibility to Bd (Scheele et al. 2017). We found relatively high
255 prevalence of Bd in some ponds of DNS but no signs of chytridiomycosis in amphibian specimens
256 and, as far as we know, no massive mortalities of amphibians have been detected in the area
257 (Hidalgo et al. 2012). Lower infection intensity of Bd found in amphibians [reaching a maximum
258 load of 68.0 ZE in one specimen of *P. walt*, which is far from the threshold of ~10,000 ZE to

259 observe mass mortalities (Vredenburg et al. 2010)], could explain why the chytrid fungus is present
260 in the area since long without relevant mass mortalities (Hidalgo et al. 2012). Another possible
261 factor is related to the virulence of Bd, which depends on each strain (Rosenblum et al. 2013);
262 however, nothing is known on which chytrid fungus strain is present in DNS.

263 Although some factors affecting Bd dynamics have been well studied (see paragraph above),
264 how the fungus can be maintained in the environment remains less known. Some amphibian
265 species, or some individuals within a species, may function as reservoirs of Bd since they are not
266 highly susceptible to infection in spite of harbouring the pathogen and transmitting it to others (van
267 Rooij et al. 2015; Scheele et al. 2017; Brannelly et al. 2018). But other studies in addition to ours
268 have demonstrated infections of Bd in non-amphibian taxa such as reptiles (Kilburn et al. 2011),
269 waterfowl (Garmyn et al. 2012; Johnson and Speare 2005), fish (Liew et al. 2017) and crayfish
270 (McMahon et al. 2013; Brannelly et al. 2015a). The generalist strategy of the Bd fungus, which is
271 able to infect a wide range of hosts, may have profound evolutionary consequences for the
272 pathogen, including the evolution of highly pathogenic strains.

273 We show that the presence of the highly invasive red swamp crayfish could be a relevant
274 factor in the prevalence of Bd in amphibians. However, our models for the infection intensity of Bd
275 in amphibians showed that crayfish presence had no effect. Thus, other variables such as amphibian
276 densities or environmental determinants (probably the hydroperiod) affecting in turn the viability of
277 Bd zoospores, could be playing a more important role in Bd infection intensity. Our study
278 demonstrates the presence of Bd in several amphibian species across DNS and a high prevalence of
279 Bd in nearby sampling sites that had been studied ten years ago (Hidalgo-Vila et al. 2012).
280 Therefore, the presence of the red swamp crayfish could explain the high prevalence of Bd, whereas
281 pond desiccation in summer could prevent sudden declines in amphibian populations in DNS. Thus,
282 human-made permanent ponds that supply water for livestock within this protected area could act
283 as infection hubs, not only because infected amphibians can remain there in summer (Hidalgo et al.
284 2012), but also because crayfish can co-occur there (Román 2014). If the red swamp crayfish acts as

285 vector competent of Bd, it would be a good step in understanding the dynamic system of this
286 disease in places where both taxa co-occur, and its participation in transmitting infection to
287 amphibians (see references in van Rooij et al. 2015).

288 The positive relationship between the presence of the red swamp crayfish and prevalence of
289 Bd in amphibians suggests an unrecognised indirect impact of crayfish on amphibian populations of
290 DNS besides the already known direct effects (Arribas et al. 2014). Although the red swamp
291 crayfish has been mainly translocated by human-mediated dispersal (Oficialdegui et al. 2019), its
292 spread in this highly protected area seems to be due to the presence of seasonal streams connecting
293 the ponds and human-made permanent ponds, thus facilitating the crayfish expansion (Román
294 2014). However, their overland movement in rice fields often depends on water availability, relative
295 humidity, air temperature and time of the day (Ramalho & Anastácio 2015), so we can expect such
296 behaviour in our study area, where temporary ponds are close to each other. Therefore, the
297 particular system of DNS, which encompasses > 3,000 temporary waterbodies, could play a crucial
298 role in the spread of the red swamp crayfish and, consequently, in the dynamics of Bd. More effort
299 should be paid in controlling emerging diseases such as Bd and the factors influencing their
300 dynamics in protected areas that are considered biodiversity hotspots.

301 To conclude, our study suggests a role of the red swamp crayfish as potential reservoir of
302 the Bd pathogen and suitable predictor of its prevalence in amphibians. The suitability of the red
303 swamp crayfish as a potential Bd reservoir (McMahon et al. 2013), together with its ubiquitous
304 presence in temperate habitats worldwide (see Oficialdegui et al. 2019) and its positive relationship
305 with high prevalence of Bd in amphibians, may represent new pathways of introduction for this
306 disease globally. This study highlights the need to include non-amphibian hosts, especially invasive
307 species such as the red swamp crayfish, as part of the dynamics of Bd infection in order to predict
308 possible outbreaks of chytridiomycosis, as well as its contemplation in future conservation
309 strategies on amphibians. Further research should consider experiments on the role of the red
310 swamp crayfish in the transmission of Bd to amphibians, longitudinal studies to assess the effect of

311 hydroperiod together with the presence of the red swamp crayfish simultaneously, as well as the
312 inclusion of non-amphibian hosts in epidemiological models to estimate clinical outcomes of this
313 panzootic disease.

314

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451 **Table 1.** The red swamp crayfish sampling in streams and ponds of western Andalusia, Spain, in
452 2015. For each location we show the coordinates, altitude (m), number of crayfish captured (n),
453 type of habitat, prevalence of Bd (%) with 95% confidence intervals in parentheses, and infection
454 intensity (Bd load in Zoospore Equivalents, mean \pm SE).

Site	Code	Name	Lat	Long	Altitude	Habitat	n	Prevalence (CI)	Bd load
1	GDP	Guadalporcún	36.948	-5.3578	282	Stream	20	0	-
2	GUA	Guadamar	37.657	-6.2290	203	Stream	20	0	-
3	HUE	Hueznar	37.933	-5.6971	422	Stream	20	0	-
4	MAN	Manecorro	37.124	-6.489	6	Lagoon	25	0	-
5	MAR	Martinazo	37.028	-6.438	8	Pond	20	0	-
6	STG	Guaperal	37.098	-6.4628	7	Pond	20	0	-
7	HOR	Hornueco	37.585	-6.5484	134	Stream	20	0	-
8	JAR	Jarrama	37.707	-6.4798	288	Stream	20	0	-
9	OLI	Olivargas	37.786	-6.8155	236	Stream	20	20.0 (5.7 – 43.7)	4.1 \pm 1.2
10	VIL	Villar	37.688	-6.7254	229	Stream	19	0	-
11	VLV	Valverde	37.568	-6.7384	233	Stream	16	0	-
12	PAL	Palmones	36.286	-5.5949	122	Stream	20	0	-
13	VAL	Valle	36.084	-5.6933	58	Stream	20	0	-
14	YES	Yeso	36.371	-5.8415	22	Stream	7	0	-
							267	1.5 (0.4 – 3.8)	4.1 \pm 1.2

Table 2. The red swamp crayfish and amphibian sampling in six ponds of Doñana Natural Space (south-western Spain) in 2018. For each location we show the coordinates, altitude (m), number of crayfish and amphibians captured (n), prevalence of Bd (%) with 95% Confidence Intervals in parentheses and infection intensity (Bd load in Zoospore Equivalents, mean \pm SE).

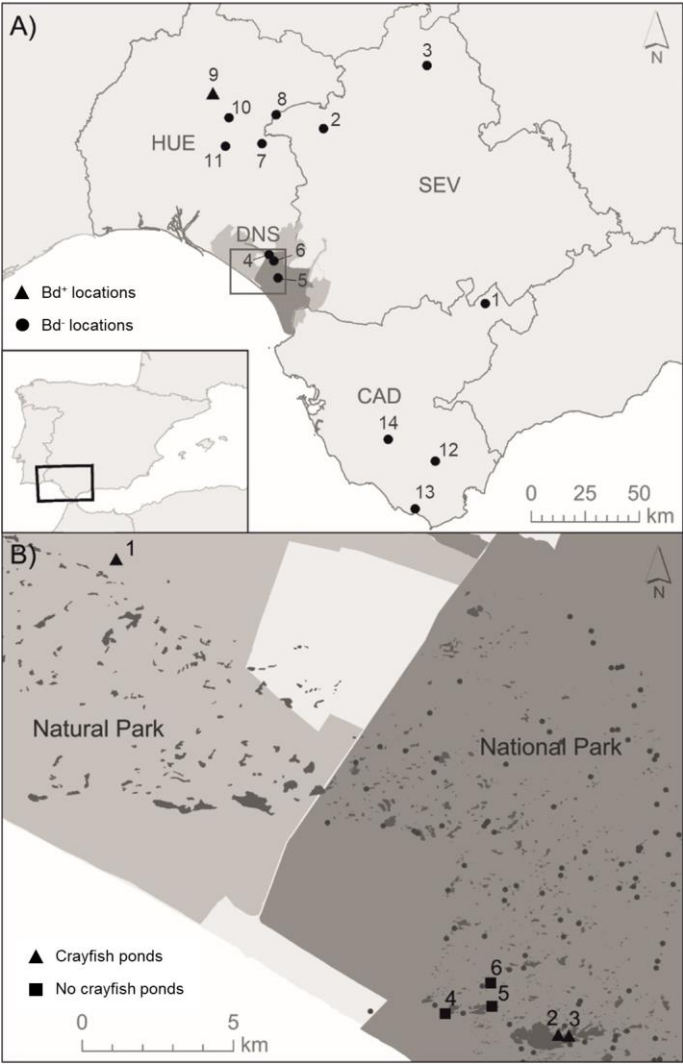
Site	Code	Location	Lat	Lon	Altitude	<i>P. clarkii</i> presence	Crayfish			Amphibians		
							n	Prevalence (CI)	Bd load	n	Prevalence (CI)	Bd load
1	ANS	Ánsares	37.122	-6.6046	23	yes	9	0.0 (0 – 33.6)	-	5	20.0 (0.5 – 71.6)	0.1 \pm 0.0
2	OLL	Santa Olalla	36.981	-6.4731	9	yes	19	26.3 (9.1 – 51.2)	7.7 \pm 2.3	40	57.5 (40.9 – 73.0)	9.3 \pm 2.3
3	PAJ	Las Pajas	36.980	-6.4699	9	yes	9	22.2 (2.8 – 60.0)	1.7 \pm 0.0	49	32.7 (19.9 – 47.5)	5.2 \pm 4.2
4	TAR	Taraje	36.989	-6.4928	10	no	-	-	-	30	10.0 (2.1 – 26.5)	0.1 \pm 0.0
5	ZAH	Zahillo	36.987	-6.5067	11	no	-	-	-	20	10.0 (1.2 – 31.7)	0.2 \pm 0.1
6	PDR	P. del Raposo	36.996	-6.4932	8	no	-	-	-	21	4.8 (0.1 – 23.8)	0.1 \pm 0.0
							37	18.9 (8.0 – 35.2)	6.0 \pm 2.0	165	27.9 (21.2 – 35.4)	1.9 \pm 0.6

Table 3. The seven amphibian species sampled in Doñana Natural Space in 2018. Life stage is represented by juveniles (J) and adults (A). The grey area includes urodelan species (salamanders) whereas the white area includes anuran species (frogs). Prevalence of Bd (%) with 95% Confidence Intervals in parentheses and Bd loads on amphibians (Zoospore Equivalents, mean \pm SE) are shown.

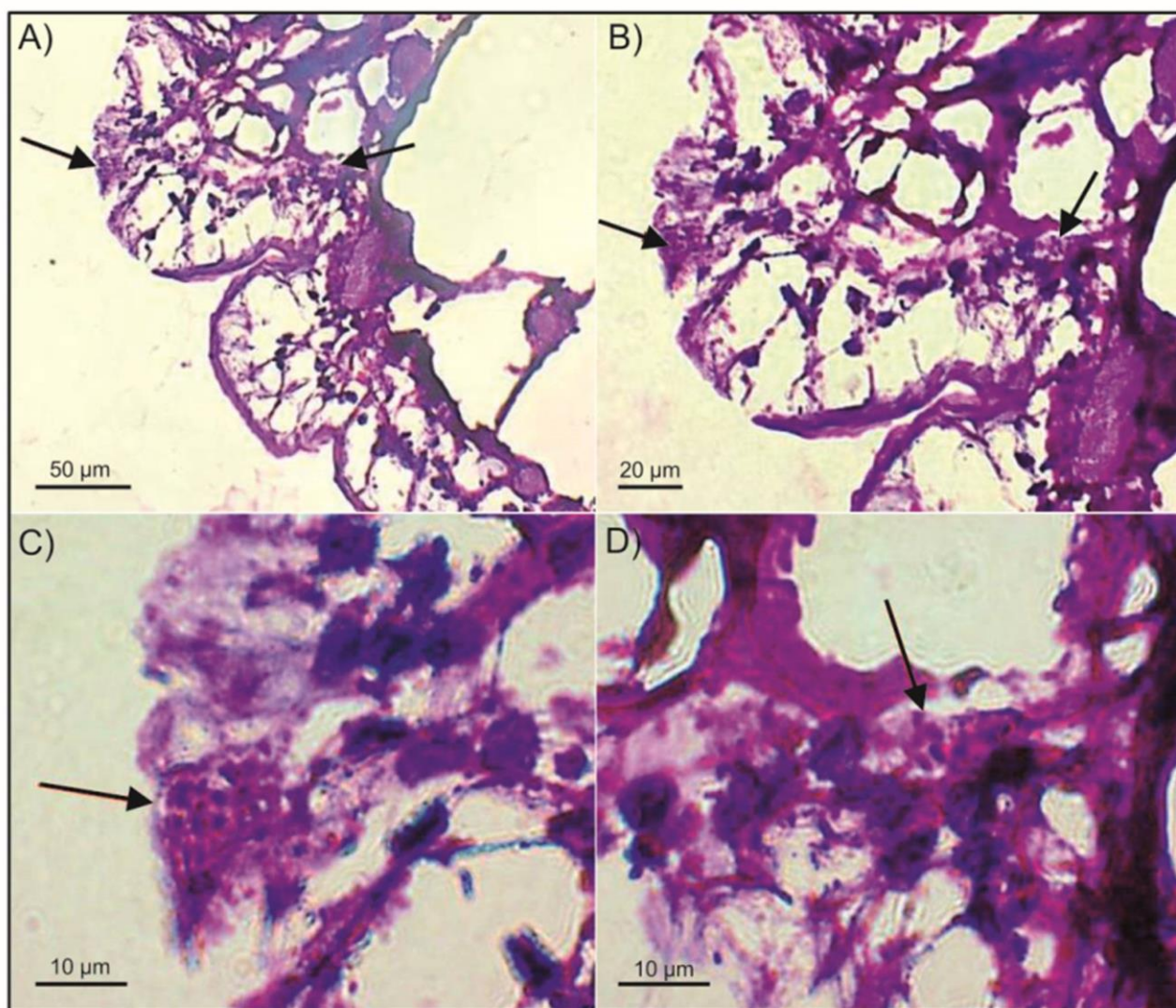
<i>Species</i>	n	Life Stage (J/A)	Prevalence (CI)	Bd load
<i>Pleurodeles waltl</i>	46	29/17	39.1 (25.1 – 54.6)	5.7 \pm 3.8
<i>Triturus pygmaeus</i>	77	0/77	35.1 (24.5 – 46.8)	7.2 \pm 2.1
<i>Discoglossus galganoi</i>	2	0/2	0.0 (0 – 84.2)	-
<i>Epidalea calamita</i>	4	1/3	0.0 (0 – 60.2)	-
<i>Hyla meridionalis</i>	29	0/29	0.0 (0.0 – 11.9)	-
<i>Pelobates cultripes</i>	5	3/2	0.0 (0 – 52.2)	-
<i>Pelophylax perezi</i>	2	2/0	50.0 (1.3 – 98.7)	0.1

467 **Figure Captions**

468 Fig. 1 (A) Map of the study area showing crayfish sampling points in streams of three provinces in
469 western Andalusia (“HUE” Huelva, “SEV” Seville and “CAD” Cadiz). Black spots indicate
470 locations with Bd- and the black triangle shows the only Bd+ location (see sites in Table 1). Area
471 depicted by a square indicates the localized highly protected area of Doñana Natural Space (DNS)
472 and greys represent the Doñana Natural Park (light grey) and Doñana National Park (dark grey). (B)
473 Protected areas of Doñana Natural Space (Doñana Natural Park and Doñana National Park).
474 Sampling sites with crayfish (black squares) and ponds without crayfish (black triangles) are
475 numerated as follows: (1) ANS, (2) OLL, (3) PAJ, (4) TAR, (5) ZAH, and (6) PDR (see Table 2 for
476 details). Dark grey patches depict temporary ponds and grey circles depict permanent water bodies.

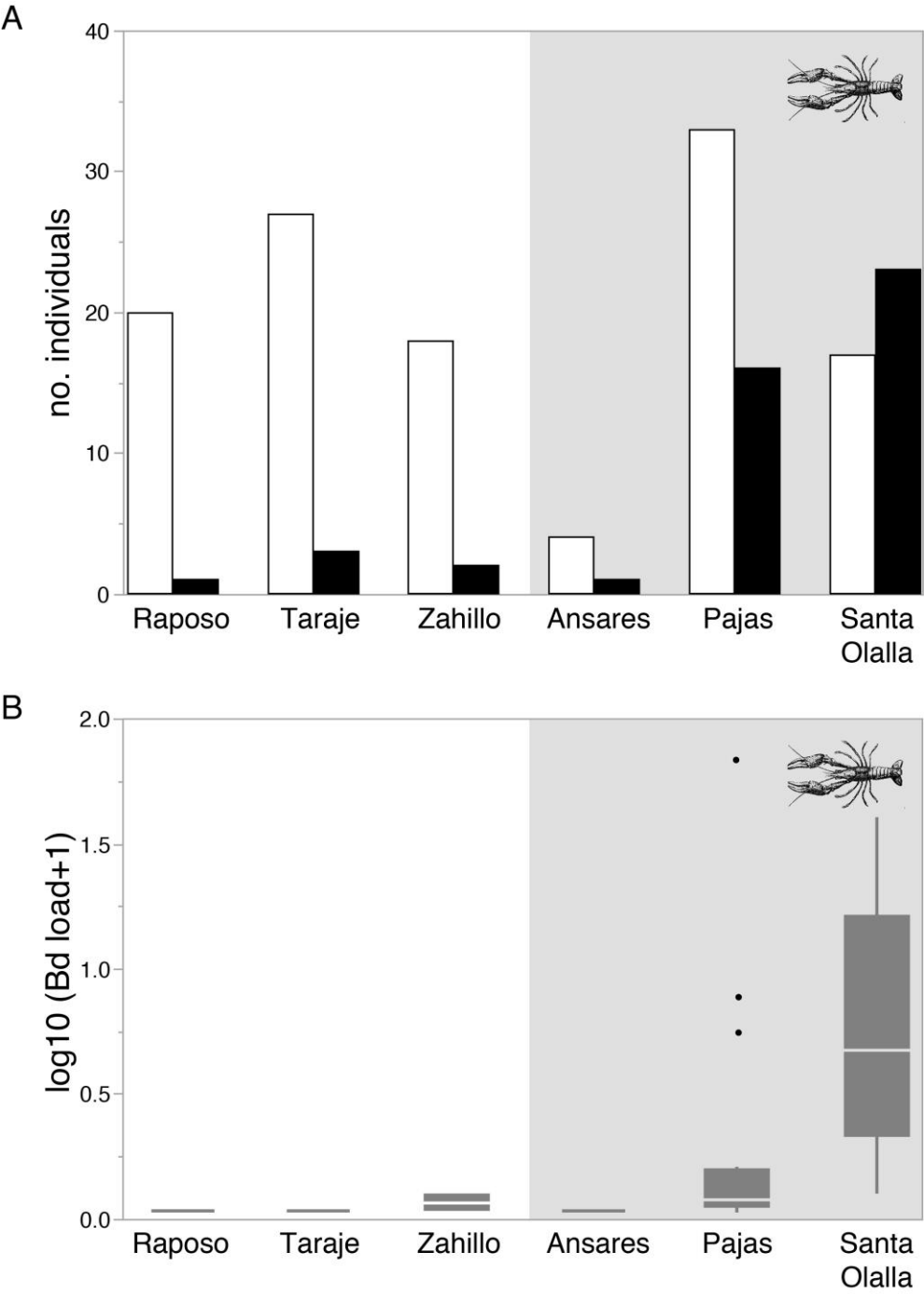


478 Fig. 2 Section of GI tissue from one adult Bd+ red swamp crayfish, *Procambarus clarkii*, at different
479 scales. Black arrows point out two different zoosporangia containing zoospores of Bd. A) The two
480 zoosporangia at 50 μm ; B) the two zoosporangia at 20 μm ; C) one of the two zoosporangia at 10
481 μm ; and D) the other zoosporangium at 10 μm .



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486 Fig. 3 (A) Prevalence and (B) infection intensity of Bd in amphibian species from six sampled
487 ponds in Doñana Natural Space. The grey area includes ponds with presence of the red swamp
488 crayfish. In figure 3A, while the number of Bd-negative are depicted in white bars, Bd-positive
489 amphibians are depicted in black bars.



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